

1 Use of LDH- chromate adsorption co-product as an air purification photocatalyst

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23 **Abstract**

24

25 This work deals with the use of layered double hydroxides for a double environmental
26 remediation. The residue obtained in the use of these materials as a chromate sorbent in water,
27 was subsequently studied as a photocatalyst for the removal of NO_x gases. With this aim, MgAl-
28 CO₃ layered double hydroxides were synthesized by the coprecipitation method with a
29 divalent/trivalent metal ratio of 3. After its calcination at 450 °C, the mixed oxide was obtained
30 and MgAl-CrO₄ were synthesized by the reconstruction method. A complete chemical,
31 morphological and photochemical study of the samples was carried out with techniques such as
32 XRD, FT-IR, TGA, XRF, PL, DRIFTS and UV-Vis spectroscopy. Results showed that LDH materials
33 presented no significant changes in their structure after their use as a sorbent. Photocatalytic
34 tests of the samples showed a very good NO removal efficiency, as well as a high selectivity (low
35 NO₂ emissions) through complete oxidation of these oxides to nitrate. The incorporation of
36 chromate into the LDH structure improved the absorption of light in the visible region of the
37 spectra, producing an improvement of 20 % in the NO elimination compared with the LDH
38 without chromate.

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40 **Keywords:** LDH, Hydrotalcite, Chromate, Photocatalyst, Nitrogen Oxides

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45 1. Introduction

46 Layered double hydroxides (LDH), also known as hydrotalcite like materials, are an important
47 class of lamellar anionic clays, which have been attracting scientific attention in recent decades.
48 Due to their unique properties and easy and inexpensive synthesis, these materials have been
49 widely studied for many applications in different fields such as catalysis (Olszówka et al., 2018;
50 Zhou et al., 2017), pharmacy (Acharya et al., 2019; Kim et al., 2020), polymer stabilizers (Taviot-
51 Guého et al., 2018), flame control (Jin et al., 2020), electrochemistry (González et al., 2016)
52 photocatalysis (Mantilla et al., 2010; Parida et al., 2012) etc. They have a brucite-like structure
53 where divalent metal ions are partially substituted by trivalent ones, and a positive charge thus
54 generated is balanced by the arrangement of hydrated anions in the interlayer (Cavani et al.,
55 1991; Rives, 2001). The composition of LDH may be expressed by the general formula $[M^{II}_{1-x}M^{III}_x$
56 $(OH)_2]^{x+} [A^{n-}]_{x/n} \cdot mH_2O$ where $M^{II} = Mg, Zn, Ni, \text{ etc.}$ $M^{III} = Al, Cr, \text{ etc.}$, $A = CO_3^{2-}, NO_3^-, Cl^- \text{ etc.}$, x is
57 the value that determines the layer charge density and the anion exchange capacity, and ranges
58 between 0.2 – 0.4. One of the peculiarities of these compounds is when calcined at 500 °C, they
59 are converted into mixed oxides, which recover the LDH structure in the presence of anions in
60 aqueous solutions (“memory effect”).

61 These versatile materials can be considered as doped semiconductors where the doping
62 agent is the transition metal cation located at the octahedral sites of the brucite sheets. The high
63 LDH photocatalytic activity is due to the easy mobility and separation of electrons and holes
64 created in these materials (Parida and Mohapatra, 2012), as well as to the surface OH^- groups
65 which can react with holes in the valence band and produce hydroxyl radicals, a key intermediary
66 in photo-oxidative reactions (Lv et al., 2020).

67 We recently reported excellent LDH photocatalytic activity for the degradation of NO_x gases
68 (DeNO_x process) (Huo et al., 2019a; Rodriguez-Rivas et al., 2018). These gases (NO_x = NO + NO₂),
69 emitted from industrial and traffic sources, are classified as priority air pollutants in urban zones
70 (European Environment Agency, 2019). They participate in the formation of tropospheric ozone
71 which, when inhaled, produce respiratory problems, reduce lung function and in general, many
72 early deaths are associated with NO_x emissions (Chen et al., 2004). Titania is the standard
73 photocatalyst used for DeNO_x applications, but it is only active under UV light ($\lambda < 387$ nm; 3.2
74 eV energy band gap) and not under visible light irradiation. Moreover, TiO₂ exhibits a low DeNO_x
75 selectivity. If we consider NO → NO₂⁻ → NO₂ → NO₃⁻ as the basic photocatalytic mechanism, a
76 huge emission of NO₂, a gas much more toxic than NO, occurs during the process. However, LDH
77 materials, besides high photocatalytic activity, also showed an outstanding selectivity (no NO₂
78 emissions) for the DeNO_x process. Thus, LDH could be a highly efficient alternative to TiO₂ in the
79 battle for NO_x abatement.

80 On the other hand, one of the widely studied LDH applications is as adsorbents for water
81 pollutants. Due to the exchangeability of interlayer anions and, the so called memory effect of its
82 calcined product, LDH have attracted great interest for removing many inorganic and organic
83 anions from contaminated waters (Chaara et al., 2012; Otero et al., 2013; Paikaray et al., 2013;
84 Pavlovic et al., 2013). The chromate ion is a very harmful water pollutant due to its mutagenicity,
85 genotoxicity, bioaccumulation and carcinogenicity. It can reach surface and groundwaters as an
86 effluent of different industries such as chrome plating, fertilizers, batteries, leather, textile, etc
87 (Fei and Liu, 2016; Kaprara et al., 2015; Koleli and Demir, 2016), as well as from filtration from
88 metal mines or inadequate handling of mining supplies (Ukhurebor et al., 2021). One of the
89 proposed methods for chromate removal from waters is through its adsorption on LDH and its

90 calcined product, and there are many reports claiming its high adsorption capacity for this
91 pollutant (Chao et al., 2018; del Arco et al., 2006; Hsu et al., 2019; Lazaridis et al., 2004; Lu et al.,
92 2016; Prasanna et al., 2006; Tran et al., 2019). However the desorption of adsorbed
93 contaminants and the regeneration of adsorbents is the most difficult and expensive part of an
94 adsorption technology, which could elevate the total cost for an adsorption system up to 70 %
95 (Goh et al., 2008).

96 The purpose of this study is to assess the reuse of the MgAl-CrO₄ adsorption residue as a
97 desirable good in the field of photocatalysis. Thus, this work deals a double environmental
98 remediation technology: removal of pollutants from water (chromate ion) and from air (NO_x
99 gases) by using a single starting component, calcined MgAl-LDH. In a first step, the MgAl LDH is
100 used to remove chromate ions from water. Secondly, once the MgAl-CrO₄ was collected and
101 dried, this co-product was reused for the photocatalytic NO_x removal from air. For this last
102 process, an enhanced response is expected due to the presence of interlayer chromate ions, in
103 similarity to that found for the chemistry parent compound molybdate which electronic
104 structure assists to the formation of electron/hole pairs (Mohapatra et al., 2012). The structural,
105 morphological and optical properties of LDH chromate adsorbate are fully characterized. The
106 changes observed are commented on with relation to their ability to perform photocatalytic
107 DeNO_x processes. The chromate containing sample shows a high photochemical activity, the
108 oxidation mechanism being explained in the light of the results obtained by using EPR, PL and
109 DRIFT techniques.

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112 2. Methods and materials

113 2.1 Synthesis of the photocatalyst

114 The preparation of $\text{Mg}_3\text{Al-CrO}_4$ proceeded in two steps: firstly, a $\text{Mg}_3\text{Al-CO}_3$ LDH ("MgAl"
115 sample) was prepared by the co-precipitation method at a constant pH = 10, at room
116 temperature (Comelli et al., 2013). Then, 0.75 mol of $\text{Mg}(\text{NO}_3)_2 \cdot 6\text{H}_2\text{O}$ and 0.25 mol of
117 $\text{Al}(\text{NO}_3)_3 \cdot 9\text{H}_2\text{O}$ mixed solution (250 mL) were added dropwise to 500 mL of magnetically stirred
118 solution containing 1.7 mol of NaOH and 0.5 mol of Na_2CO_3 . The suspension thus obtained was
119 aged for 24 h at room temperature. Then, it was centrifuged and washed several times with
120 distilled water to eliminate the presence of nitrates, and dried. The resulting product was then
121 thermally treated in a furnace at 500 °C for 6 hours. 2 g of thus obtained Mg-Al mixed oxide
122 (named MgAl/500) was added to a 100 mL of 0.05 M K_2CrO_4 solution and stirred during 24 h. to
123 obtain reconstructed $\text{Mg}_3\text{Al-CrO}_4$ LDH (named MgAl-Cr). Then, it was centrifuged, washed
124 several times and dried for 24 h at 90 °C.

125

126 2.2 Characterization of the samples

127 X-ray diffraction powder (XRD) patterns were obtained using a Bruker D8 Discover
128 instrument, with Cu K α radiation ($\lambda = 1.5406 \text{ \AA}$). Infrared spectra (IR) analyses were carried out
129 on transmission mode in a FT-MIR Perkin Elmer FRONTIER with a resolution of 4 cm^{-1} . Chemical
130 analyses were determined by X-ray fluorescence (XRF) on an ZSX Primus IV Rigaku. The water
131 content of LDH was calculated from the thermogravimetric (TG) analysis which was obtained by
132 using a Mettler Toledo apparatus in air atmosphere (flow: 100 mL min^{-1}) at a heating rate of 5 °C
133 min^{-1} . Nitrogen adsorption–desorption isotherms were recorded in an ASAP 2020 apparatus
134 (Micromeritics). Prior to the sorption measurements, samples were degassed at 100 °C under
135 vacuum for 4 h. The morphological characterizations were done by scanning electron

136 microscopy (SEM) using a Jeol JSM 7800F microscope. Diffuse reflectance UV-Vis (DRS) spectra
137 of the samples were recorded from 200 to 800 nm employing a Cary 5000 spectrophotometer
138 with PTFE as reference.

139

140 *2.3. Photocatalytic activity evaluation*

141 Photocatalytic activity was examined analyzing the removal ratio of NO in a continuous-flow
142 reactor illuminated by a sunlight irradiation box (Solarbox 3000e RH; Xe lamp with irradiances of
143 25 and 580 W·m⁻² for UV and visible light respectively). The reactor contained the LDH
144 photocatalyst. 300 mg of sample powder, supported in a 5x5 cm sample holder, was used in
145 each experiment. 500 ppb NO simulated polluted atmosphere flowed through the reactor (0.3 L
146 min⁻¹) with a relative humidity 50 ± 10 %. Firstly, the adsorption-desorption equilibrium between
147 gas and sample was assured for 10 min. Subsequently, the lamp was switched on for 30 min. The
148 concentration of NO was continuously measured by a NO_x chemiluminescence analyzer
149 (Environnement-AC32M).

150 Electron paramagnetic resonance (EPR) technique assisted in corroborating the radical
151 species involved in the photocatalytic process. The spectra were recorded at room temperature
152 in a Bruker-EMX micro spectrometer. The spin-trap 5,5-Dimethyl-1-pyrroline N-oxide (DMPO,
153 Sigma-Aldrich) was used for radical detection. Water or methanol solutions were prepared for
154 detecting ·OH or ·O₂⁻ radicals, respectively. The photocatalyst specimen was immersed in a 45
155 mM DMPO solution and irradiated for 30 min with sunlight. Photoluminescence (PL)
156 measurements were carried out in a FLS920 Fluorimeter (Edinburgh Instrument Ltd, Livingston,
157 UK). A 450 W Xe lamp was employed as the excitation source in the steady-state measurements
158 and a picosecond pulsed diode laser ($I_{exc} = 406$ nm and pulse width = 87 ps) was used for the

159 time-resolved measurements. In order to find out about the photochemical DeNOx mechanism,
160 in situ diffuse reflectance infrared Fourier transform spectroscopy (DRIFTS) measurements were
161 performed using a Perkin Elmer FRONTIER FT-IR spectrometer equipped with a Harrick reaction
162 chamber. The sample was placed in the reaction chamber, treated at room temperature under
163 high-purity Ar flow for 10 minutes to remove environmental impurities. Subsequently, in the
164 dark, a mixture of 50 mL/min NO (100 ppm) and 50 mL/min O₂ was pumped into the chamber,
165 the sample being subjected to adsorption reactions for 25 minutes. Afterwards, light was turned
166 on for 60 min. The spectra evolving in both periods were sequentially recorded every 5 minutes.

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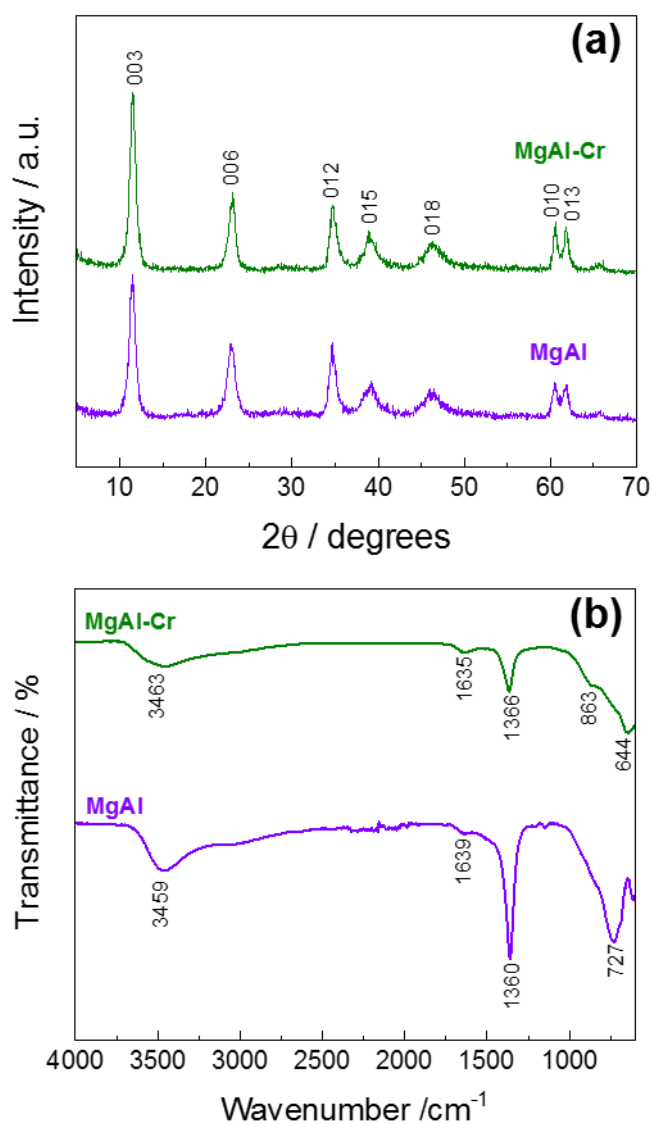
168 3. Results and discussion

169 3.1 Characterization of photocatalysts

170 The MgAl and MgAl-Cr XRD patterns (**Figure 1a**) were both identified as hydrotalcite-like
171 compounds indexed in a rhombohedral lattice (Cavani et al., 1991). MgAl-Cr shows strong (00l)
172 plane reflections, which confirms the rehydration and reconstruction of the mixed oxide
173 MgAl/500 (**Figure S1**) into the LDH, after its exposure to an aqueous solution of chromate anions.
174 The first two peaks at 10-11°/2θ and 22-23°/2θ correspond to the basal reflections 00l, which in
175 the case of MgAl-Cr are shifted to the right indicating a decrease in its interlayer space,
176 compared to those of MgAl from $d = 7,87 \text{ \AA}$ to $d = 7,72 \text{ \AA}$ (**Figure 1a**). The shortening of the
177 interlayer distance in these layered structures is related to the grafting of CrO₄²⁻ to the metal
178 hydroxide slab by displacing OH⁻ molecules (del Arco et al., 2004). After grafting, the interlayer
179 anions are better attached to the layers and are much more difficult to exchange with other
180 anions (Costantino et al., 1997). This is very important for the scavenging of toxics as it will
181 prevent the eventual leaching of chromate ions from the hydrotalcite interlayer. On the other

182 hand, a part of the chromate anions could be adsorbed by a surface complexation at the edges
183 of hydrotalcite particles (Jobbágy and Regazzoni, 2013). The patterns were indexed in a
184 rhombohedral lattice with a values $[a = 2d_{(110)}]$ 3.06 Å and c values 23.13 Å and 23.438 Å
185 $[c = 3d_{(003)}]$ corresponding to MgAl and MgAl-Cr, respectively. The crystal size was D=90.4nm for
186 MgAl and 90.2nm for MgAl-Cr.

187 The Fourier Transform Infrared (FT-IR) spectra of both samples are shown in **Figure 1b**, being
188 characteristic of LDH compounds. In both cases, a broad band observed around 3500 cm^{-1}
189 corresponds to the stretching vibrations of the hydrogen-bonded hydroxyl group of both
190 hydroxide layers and interlayer water molecules, while the band at $\sim 1639 \text{ cm}^{-1}$ is attributed to O-
191 H bending vibration of water molecules. The MgAl sample exhibits a strong band at 1360 cm^{-1} ,
192 which corresponds to the asymmetric stretching mode of carbonate interlayer anions. However,
193 the presence of a weaker carbonate band (1366 cm^{-1}) is also observed in MgAl-Cr because,
194 despite the precautions taken during the synthesis, some intercalation of carbonate is difficult to
195 avoid, due to its high reactivity and affinity for the hydrotalcite interlayer. Chromate
196 incorporation in the LDH structure is verified by the appearance of an absorption band at 863
197 cm^{-1} , which corresponds to the stretching frequency ν (Cr–O) (del Arco et al., 2006). Compared
198 to the free anion, this band is slightly shifted toward lower frequencies (27 cm^{-1}), indicating the
199 lowering of symmetry of the chromate ion in the LDH interlayer (Malherbe et al., 1999; Prasanna
200 et al., 2006). Absorptions below 800 cm^{-1} are due to lattice vibrations, involving metal–oxygen
201 stretching modes.



202

Figure 1. (a) XRD patterns and (b) IR spectra obtained from the MgAl and MgAl-Cr samples.

203 The chemical analysis of the original and modified MgAl-LDH is shown in **Table 1**. In both
 204 cases, the Mg:Al molar ratio is similar to that of the starting solution. **Figure S2** shows that the
 205 thermal behavior of Mg–Al LDH is characterized by two main steps: up to 250 °C the physisorbed
 206 and interlayer water is lost, while at 250–500 °C the simultaneous dehydroxylation and
 207 decarbonization processes take place and the LDH converts into a MgAl mixed-oxide (Frost et al.,
 208 2005; Rives, 2001) (**Figure S1**). The amount of interlayer water, determined by the
 209 thermogravimetric curves from the first weight loss, together with the Mg/Al molar ratio

210 obtained by XRF analysis were used to propose the chemical formula of the samples. It is
 211 inferred from these results that, based on this synthetic procedure, around 17 % of anions per
 212 LDH formula can be retained as chromate.

Table 1. *Physicochemical characterization for the LDH samples.*

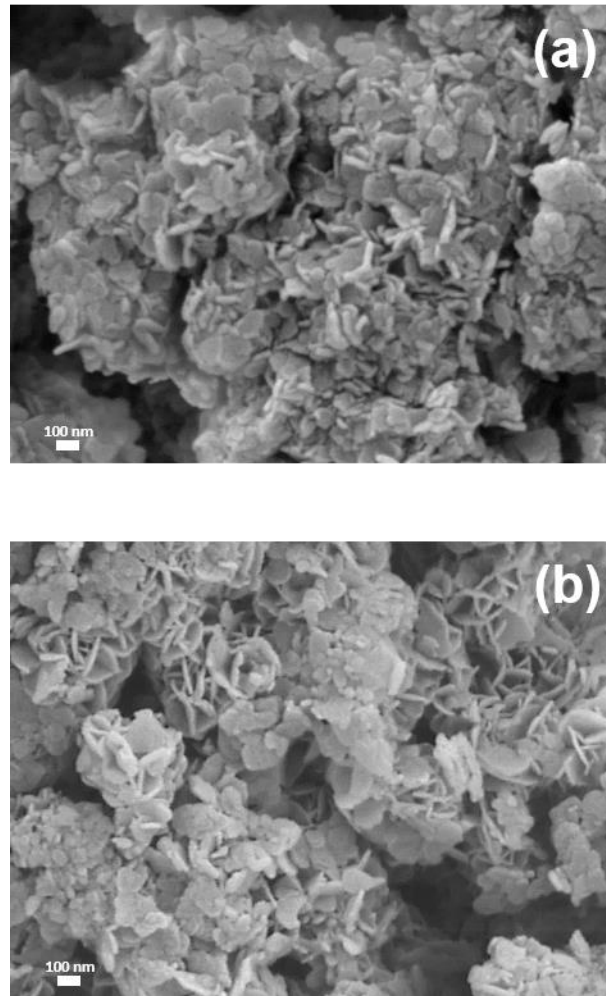
LDH	% w/w			M ²⁺ /M ³⁺ ratio	Proposed formula	S _{BET} (m ² g ⁻¹)	Lattice parameters		
	Mg	Al	Cr				D (nm)	a (Å)	c (Å)
MgAl	23.5	8.68	<LOD	3	[Mg _{0.75} Al _{0.25} (OH) ₂](CO ₃) _{0.125} ·0.62H ₂ O	30	92.4	3.06	23.13
MgAl-Cr	22.3	8.54	1.85	2.9	[Mg _{0.745} Al _{0.255} (OH) ₂](CrO ₄) _{0.036} (CO ₃) _{0.092} ·0.69H ₂ O	74	90.2	3.06	23.43

213

214 The morphology of the samples was evaluated by SEM observations, **Figure 2**. The MgAl
 215 sample appears as agglomerates of poor crystallized pseudo-hexagonal nanoparticles (**Figure**
 216 **2a**). The reconstruction of the LDH structure during the inclusion of the chromate anion leads to
 217 a better crystallization and growth of the hexagonal particles, but also to the appearance of a
 218 high presence of smaller unshaped particles (**Figure 2b**). In consequence, the pore
 219 microstructure analyzed for the samples was somewhat different. **Figure S3** shows the N₂
 220 adsorption-desorption isotherms measured for each sample. Slight differences were found on
 221 the isotherms shape. The presence of a hysteresis loop suggests the presence of mesopores,
 222 which were more pronounced in the case of the MgAl sample (from 2 to 40 nm; inset Figure
 223 S3a), fitting to a type IV isotherm. However, a large fraction of macropores was also observed in
 224 the case of the MgAl-Cr sample (from 50 to 160 nm; inset Figure S3b) exhibiting a type II shape
 225 (Rodriguez-Rivas et al., 2018), that assigned to the adsorption on macro and non-porous
 226 materials. The specific surface area values (SSA; BET method) for both samples are given in **Table**
 227 **1**. As usually accounts for reconstructed LDHs an increase on SSA is observed (Rives, 2001). Thus,
 228 a significant value of 74 m²·g⁻¹ was measured for the MgAl-Cr sample, being more than double of

229 that exhibited by the MgAl sample, this being of importance for its potential use in
230 heterogeneous photocatalytic processes.

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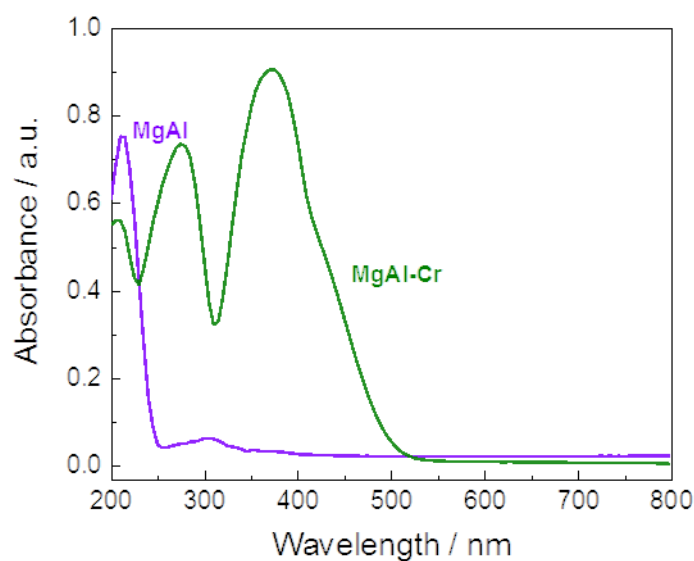
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Figure 2. SEM images of the (a) MgAl and (b) MgAl-Cr samples.

233

234 Finally, in that concerning the characterization of the samples, the light absorption spectra
235 were recorded, **Figure 3**. The spectra obtained for the MgAl sample is similar to that previously
236 reported for this system (Gao et al., 2019). Because of its large band-gap value, $E_g > 5.0$ eV (Lv et
237 al., 2020; Mantilla et al., 2011), an intense peak at 220 nm is observed in the UV light region.
238 Considering the activation of environmental photocatalytic processes by sunlight, the

239 observation of bands at $\lambda > 250$ nm is most interesting. Thus, on this sample, a small absorption
240 band at 305 nm in the UV light region, is observed, with a tail extending to the visible region until
241 430 nm. On the other hand, two intense UV bands are observed at 275 and 370 nm for the
242 chromate intercalated sample, the last showing a shoulder in the visible light region at 445 nm.
243 This spectra resembles that of chromate (VI) ions in solution and differs from those of samples
244 containing insoluble chromate (VI) (Kuncewicz et al., 2011; Zheng et al., 2018), which could
245 indicate the presence of CrO_4^{2-} ions grafted to metal hydroxide sheets. In summary, the spectra
246 have shown that the chromate containing LDH sample exhibits a higher ability to harvest sunlight
247 radiation.



248

Figure 3. UV-Vis absorption spectra of the MgAl and MgAl-Cr samples.

249 3.2 Photocatalytic behaviour in DeNO_x processes.

250 Once the chromate ion is retained in the LDH structure and the resulting LDH-chromate
251 complex shows a noticeable light absorption ability, it is interesting to test its potential use in
252 environmental remediation processes assisted by photocatalysis. Thus, the efficiency in the
253 photocatalytic NO gas removal from air was studied as an air purification technique. This process

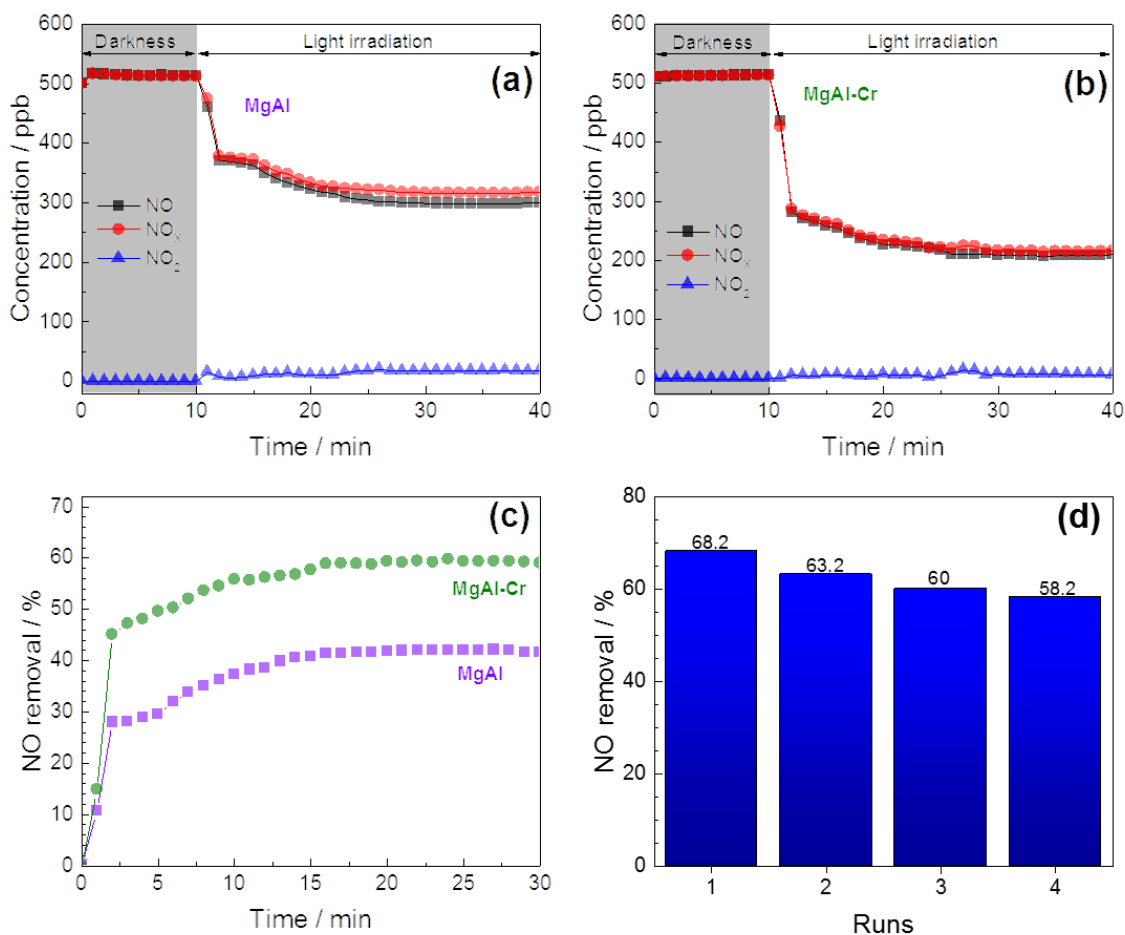
254 consists in promoting the photochemical oxidation (PCO) of nitrogen oxide gases to
255 nitrite/nitrate ($\text{NO}_2^-/\text{NO}_3^-$) species, which are removed from air and retained on the surface of
256 the photocatalyst. The successful use of different LDH as DeNOx photocatalysts has recently
257 been reported (Fragoso et al., 2021; Pastor et al., 2020; Rodriguez-Rivas et al., 2020, 2018).
258 Moreover, the MgAl-LDH system has also shown a modest activity in this process (Lv et al.,
259 2020).

260 When the LDH photocatalyst is light irradiated, the electron in the valence band (VB) can
261 acquire the necessary energy to jump to the conduction band (CB). Then, pairs of mobile charges
262 (e^- and h^+) are generated and, once they reach the surface of photocatalyst particles, react with
263 the adsorbed water and oxygen molecules. In consequence, reactive oxygen species (ROS),
264 mainly hydroxide ($\cdot\text{OH}$) and superoxide radicals ($\cdot\text{O}_2^-$) are formed, initiating a fast oxidation of
265 nitrogen oxide gases because of their powerful oxidant character. The evolution of the nitrogen
266 oxide concentration profiles in the presence of the MgAl-LDH samples were recorded with the
267 function of light irradiation time, **Figure 4**. During the first ten minutes, because there was no
268 light irradiation, the NO concentration in the reaction chamber remained constant to the inlet
269 value. However, a sudden decrease in the NO concentration values occurred once the lamp was
270 switched on, evidencing the participation of a light activated process (**Figures 4a and 4b**). For
271 both samples, the PCO process reached a stationary state after fifteen minutes of light
272 irradiation. Following these experiments, the ability of each sample to remove NO gas from air
273 could be related with the decrease in NO concentration values under light irradiation. The
274 modest NO removal efficiency of 40 % for the MgAl sample agrees with its low ability to harvest
275 sunlight ($\lambda > 250 \text{ nm}$), **Figure 4c**. However, a remarkable efficiency of 60 % was measured for the
276 chromate containing sample; a value in line with the best reported for LDH DeNOx
277 photocatalysts (Fragoso et al., 2021; Pastor et al., 2020; Rodriguez-Rivas et al., 2020). This test

278 proves that LDH chromate-sorbents can successfully be reused as air purification photocatalysts.
279 Moreover, no emission of NO₂ gas was displayed during the PCO process (**Figures 4a and 4b**).
280 NO₂ is an expected intermediate in this PCO process (Balbuena et al., 2015; Rodriguez-Rivas et
281 al., 2020), but it must be avoided since it is much more toxic than NO (Lewis, 2012). Therefore,
282 the MgAl-Cr sample is a highly efficient DeNOx photocatalyst, but also environmentally
283 sustainable because there is no release of NO₂ gas. Finally, this photocatalyst was studied in
284 successive DeNOx runs exhibiting a small loss of NO removal efficiency, **Figure 4d**.

285 The higher photocatalytic efficiency exhibited by MgAl-Cr, in comparison with MgAl, should
286 be associated to their different microstructural and electronic properties. The higher specific
287 surface area exhibited, once the LDH is reconstructed with CrO₄²⁻, allows many active sites on
288 particle surface to be available for the gaseous reactant molecules, increasing the process
289 efficiency. On the other hand, the harvesting of the sunlight is remarkably done through the
290 incorporation of chromate ions, which should eventually lead to a higher formation of ROS
291 species and better photochemical activity. It is known that the electronic structure of a LDH
292 could be influenced by different factors: the ionic nature and ratio of metal elements, as well as
293 the nature of the interlayer anion (Wu et al., 2018). On the other hand, the important role of
294 interlayer anions in the photocatalytic activity of LDHs has also been reported (Mohapatra and
295 Parida, 2016). Therefore, changes in the electronic structure of MgAl-LDH are expected after the
296 incorporation of CrO₄²⁻ ions. Because of its similarity to the molecular electronic structure, the
297 participation of chromate on the LDH electronic structure should be considered like that
298 reported for molybdate intercalated LDH (Mohapatra et al., 2012). Following the predicted
299 molecular orbital energy level diagrams for tetrahedral oxyanions of transition elements, the
300 non-bonding HOMO orbital of chromate is a fully filled oxygen 2p orbital, while the LUMO is the
301 molecular orbital contributing Cr (3d orbital). Thus, ligand to metal charge transfer from HOMO

302 to LUMO should take place under light irradiation forming additional e^- and h^+ pairs, which
303 favour a major production of ROS species.



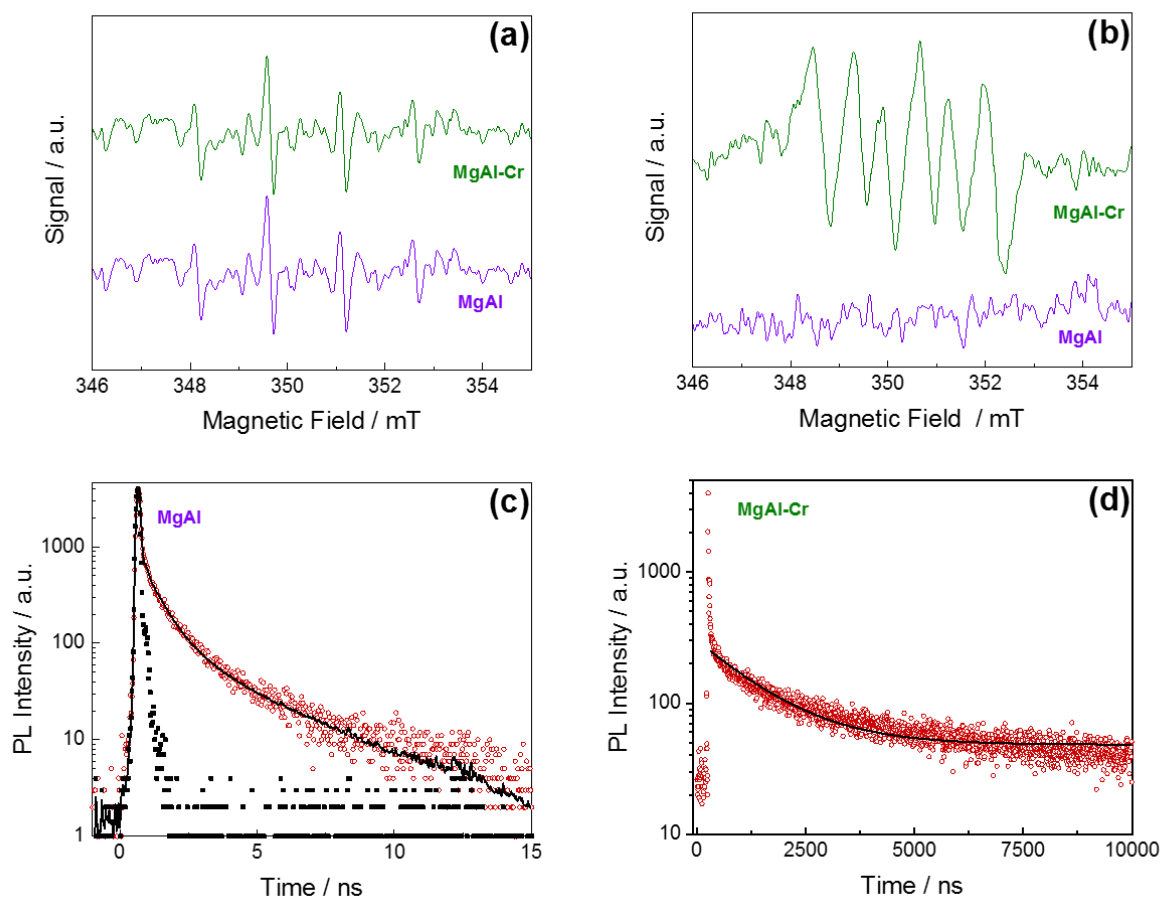
304 **Figure 4.** *NO, NO₂ and NO_x gases concentration evolution during the photo-degradation of gaseous NO under UV-Vis light irradiation on the (a) MgAl and (b) MgAl-Cr samples. (c) NO removal efficiency for both LDH samples. (d) No removal efficiency calculated for each run of the reusability tests for the MgAl-Cr sample.*

305

306 The EPR experiments are in line with the assumed electronic structure. Figure 5 shows the
307 EPR signal obtained using MgAl-LDH as photocatalysts and DMPO as the spin-trapping agent
308 under sunlight excitation, in aqueous and methanol solutions (**Figures 5a** and **5b**, respectively).
309 No signal was obtained for the photocatalyst suspension (blank) in darkness. An identical signal
310 for the DMPO- \cdot OH adduct was registered for both examples, indicating a similar ability to

311 produce electron holes. However, in the case of the DMPO- $\cdot\text{O}_2^-$ signal huge differences were
312 found, this signal being negligible for MgAl samples but very intense for MgAl-Cr. Therefore, the
313 production of excited electrons seems to be favored by the presence of chromate ions, probably
314 assisted by the injection of additional electrons from the HOMO \rightarrow LUMO transition. In line with
315 the results obtained from DeNOx tests, the clear production of superoxide radicals with the
316 MgAl-Cr photocatalyst turns in a key factor enhancing the photochemical oxidation of NO gas.

317 An efficient charge separation and transfer of the photogenerated electron-holes is essential
318 for the enhanced photocatalytic activity. To reveal the effect of chromate in the separation of
319 electron-hole pairs, PL spectroscopy is employed to clarify the ability of the recombination of
320 photogenerated species. **Figure S4** exhibits a PL band centred around 405 nm in the MgAl
321 sample ($\lambda_{\text{exc}} = 315$ nm) while a red-shifted PL peak is detected around 560 nm for the MgAl-Cr
322 sample ($\lambda_{\text{exc}} = 400$ nm). The MgAl-Cr sample was also excited at 315 nm but no PL was detected
323 which indicates a different deactivation mechanism in both materials. The shift in the PL position
324 is clear evidence of the presence of the chromate species in the MgAl-Cr sample. Thus, the
325 addition of the chromate species modifies the energetics of the electronic bands opening a
326 radiative mechanism at lower energy. In addition, **Figure 5** displays the PL decay of both samples
327 where mono- or bi-exponential functions were used to fit the experimental data. Time constant
328 of 0.67 and 3.05 ns were obtained for the MgAl sample (**Figure 5c**) in agreement with previous
329 reports, while a longer time constant of 1341 ns was obtained for the MgAl-Cr sample (**Figure**
330 **5d**). The extended time for the excited electron-hole pairs in the MgAl-Cr sample forecasts a
331 larger reaction with O_2 to generate the superoxide species, $\cdot\text{O}_2^-$, as previously observed in the
332 EPR measurements.



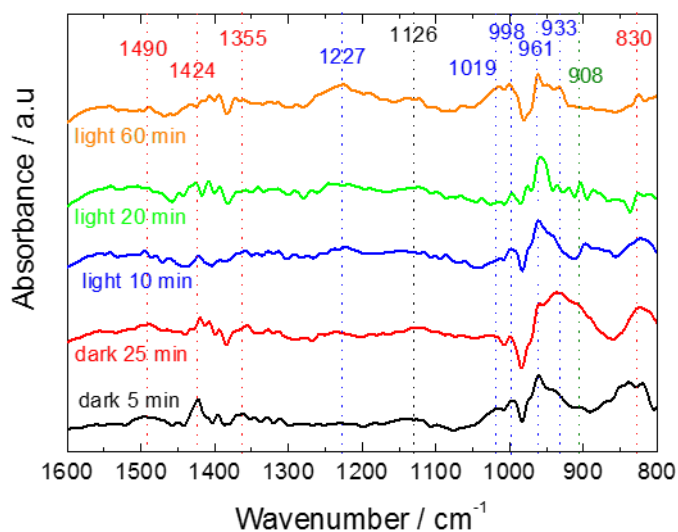
333

Figure 5. DMPO spin-trapping EPR spectra of the MgAl and MgAl-Cr samples under UV-Vis light irradiation for 15 min in (a) aqueous solution for $\cdot\text{OH}$ and (b) methanol solution for $\cdot\text{O}_2^-$. Decay times of the (c) MgAl and (d) MgAl-Cr samples.

334

335 Finally, with the aim of discerning the photocatalytic removal mechanism of NO taking place
 336 on the sample surface, in situ DRIFTS measurements were performed. Firstly, the NO adsorption
 337 processes in the absence of light irradiation were analyzed. **Figure 6**, shows sequential spectra
 338 obtained in dark conditions under NO flow for the MgAl-Cr sample, once the background
 339 (spectrum obtained before NO flow) had been subtracted. During the first 5 minutes, signals
 340 corresponding to nitrogen oxide species were registered. The appearance of NO^- (1126 cm^{-1}) and
 341 NO_2^- species ($830, 1355, 1424$ and 1490 cm^{-1}) (Huo et al., 2019b; Li et al., 2019; Liao et al., 2020;
 342 Zhang et al., 2018) is a consequence of the NO disproportionation in the presence of the

343 hydroxyl LDH groups (Liao et al., 2020); therefore the adsorption of NO occurs on this
344 photocatalyst. These species were ultimately oxidized to nitrates (933, 961, 998 and 1019 cm^{-1}
345 ¹)(Zhang et al., 2018). After 25 minutes, as higher amounts of NO came into contact with the
346 surface, they also suffered oxidation to N_2O_4 , signal at 908 cm^{-1} (Chen et al., 2020).
347 Subsequently, the sample was light irradiated and then a sequential disappearance of lower
348 oxidation nitrogen oxides species was observed, being photochemically oxidized to nitrate (961,
349 998 and 1019 cm^{-1}). The latter was the most abundant specie on the surface, the monodentate
350 nitrate being gradually transformed with time into more stable bridged linkage (1227 cm^{-1})
351 (Chen et al., 2020). Moreover, during the irradiation period signals of NO_2 or N_2O_4 were not
352 observed, in line with the high DeNOx selectivity shown by this photocatalyst.



353

Figure 6. DRIFTS spectra of MgAl-Cr, NO adsorption un darkness and NO oxidation processes under light illumination.

354

355 4. Conclusions

356 Mixed oxide obtained by calcination of MgAl LDH is proved as an efficient sorbent to capture
357 chromate ions from aqueous solutions. Compared to the pristine MgAl-CO_3 LDH, the

358 reconstructed LDH containing chromate exhibits a modified morphology, different micropore
359 structure and higher specific surface. Moreover, the presence of CrO_4^{2-} anions enhances the
360 ability of the LDH sample to absorb UV and Visible light. These characteristics promote the use of
361 MgAl-Cr as a promising photocatalyst. In fact, good efficiency for NO removal is measured, which
362 is oxidized towards nitrate on the photocatalyst surface under light irradiation. Interestingly,
363 MgAl-Cr performs as an ecofriendly photocatalyst as no emissions of toxic NO_2 are accounted for
364 during the PCO process. EPR and PL results highlight the role of chromate in the enhanced
365 photoactivity, facilitating the production of photoexcited electrons and retarding the electron-
366 hole recombination. The obtained results point out that LDH co-products obtained from water
367 depollution tasks are valuable for additional environmental processes, the air purification
368 through photocatalysis in this case, opening an interesting field of study in environmental
369 chemical engineering.

370

371 **Credit author statement**

372 **A.Nehdi:** Investigation **N.Frini-Srasra:** Funding acquisition, Supervision **G. de Miguel:**
373 Investigation, Writing and Reviewing. **Pavlovic:** Conceptualization, Writing, Review and Editing;
374 Visualization **L. Sánchez:** Conceptualization, Writing, Review and Editing, Project administration
375 **J. Fragoso:** Investigation, Validation, Writing, Review and Editing, Visualization

376

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