

Applied Vegetation Science**Modelling *Buxus balearica* distribution in Southern Spain.**

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1 **Modelling *Buxus balearica* distribution in Southern Spain.**

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14

15 **Abstract**

16 **Question:** A comparison between predictive models to study the habitat distribution of *Buxus*
17 *balearica* Lam. at a local scale.

18 **Location:** The distribution area of *Buxus balearica* studied was the Málaga and Granada
19 populations (Southern Spain), which includes almost its total Iberian Peninsula distribution
20 (36° and 38° N). The study area was restricted to basic soils, mainly marls and chalky ones,
21 which occupied a surface of 38,180 km².

22 **Methods:** The prediction models were tested based on 17 environmental variables. Six
23 methods were compared: Multivariate adaptative regression spline (MARS), Maximum
24 entropy approach to modelling species' distributions (Maxent), two generic algorithms based
25 on environmental metrics dissimilarity (Bioclim and Domain), Algorithm for Rule-set
26 Prediction (GARP), and supervised learning methods based on generalized linear classifiers
27 (Support vector machines-SVMs). As a test of the predictive power of the models we used
28 *Kappa* index.

29 **Results:** Maxent showed the better predictive accuracy, following by GARP models. All the
30 other models tested obtained lower accuracy values. By comparing the predictive power of
31 Models, climate variables showed the highest contributions of the environmental variables
32 studied. The variables with lowest contributions were the isolation models. A test of
33 sensitivity to the reduction in the number of variables obtained an accuracy of over 0.90 by
34 applying just 3 climatic variables (spring rainfall, mean temperature of warmest month, and
35 mean temperature of coldest month). All algorithms produced maps that coincided well with
36 the known distribution of the species.

37 **Conclusions:** Maxent has showed the better habitat-specific predictive mapping model of the
38 *Buxus balearica* communities in southern Spain.

39 **Keywords:** Environmental gradient; Habitat distribution model; Mediterranean shrubs.

40 1. Introduction

41 The genus *Buxus* is represented in Spain by two species, *B. sempervirens* L. and *B.*
42 *balearica* Lam. *B. balearica* is an evergreen shrub, rarely small trees, monoecious, more
43 robust and bigger than *Buxus sempervirens*, up to 5 m. high. It has glabrous stems and entire,
44 coriaceous and opposite (9-18 x 25-40 mm) leaves, dark green on their adaxial side and light
45 green on their back, with short petiole. Small flowers in inflorescence c. 10 mm in some leaf
46 axils, where only one is female, sessile in the centre of the clusters, and the rest are male,
47 shortly pedicelled with four yellowish pieces and four stamens; the female ones have three
48 persistent styles. Flowering is IV-VI, and fruits are ovoid, loculicide capsules, trilocule, up to
49 7 mm. widely oblong, with its styles as long as its capsule. Seeds are black, with 2 in each
50 locus.

51 Chorological studies show that *B. balearica* is a relict of Tertiary flora, which was
52 much more abundant before but is nowadays relegated to calcareous soils at sea level or close
53 to the coast. *B. balearica* is confined to the Western Mediterranean area, concretely in the
54 Balearic Islands, Sardinia, and North of Morocco and also in sub-coastal mountains in the
55 South East of the Iberian Peninsula (Lázaro et al. 2006). Two other populations are also
56 known in Turkey, initially considered a different species. Its current distribution in Andalucía
57 consists of a few fragmented and isolated populations distributed throughout it from Malaga
58 to Almeria (Blanca 1999). Its populations have drastically decreased and a significantly
59 reduced number of individuals are found due to both climate change in the Mediterranean
60 region during the Holocene (Yll et al. 1997) and human activity during the last centuries (Yll
61 et al. 1997, Lázaro et al. 2006). It has been catalogued as “vulnerable” by the UICN, and in
62 Spain, it is considered to be at risk of extinction in Andalusia (Blanca 1999).

63 All the evidence shows that *B. balearica* in southern Spain finds acceptable conditions
64 for flowering; fructification, seed production, dispersion and germination (Lázaro et al. 2006).

65 However, summer drought, the most crucial factor affecting the regeneration process in a
66 number of Mediterranean species promotes an extreme environmental restriction for seedling
67 establishment. Populations are thus doomed to disappear in the long term. The situation is
68 even tougher due to the effect of wild fires and human pressure, which accelerate the
69 population's fragmentation process and vanishment. This situation is creating a restrictive
70 scenario for the species' conservation programs, demanding a higher precision in repopulation
71 activities.

72 Ecological niche modelling is important for a variety of applications in ecology and
73 conservation (Graham and Hijmans 2006). For example, they attempt to provide detailed
74 distribution cartography related to the spread of species (Thuiller et al. 2004), impacts on
75 climate change (Thomas et al. 2004, Matsui et al. 2009), and spatial patterns of species
76 diversity (Feldemeyer-Christe et al. 2007). As such, ecological niche modelling has been the
77 target for an impressive growth in attention in recent years (Guisan & Zimmermann 2000,
78 Thuiller et al. 2004, Thomas et al. 2004, Araújo et al. 2005, Soberón & Peterson 2005,
79 Graham and Hijmans 2006; Elith et al. 2006, Matsui et al. 2009) placing this technique among
80 emerging new approaches relevant to ecology, biogeography, and conservation biology.

81 The basic approach of these models is to combine a set of known occurrences together
82 with prediction variables such as topographic, climatic, edaphic, biogeography and remotely
83 sensed ones. Accurate occurrence data (presence and absence) are rarely available, especially
84 for rare species or inaccessible site locations. Correlative models using species presence and
85 absence locations records for habitat predictions have been referred to as discrimination
86 techniques, while those using only species presence records have been referred to as profile
87 techniques (Jeschke & Strayer, 2008). Examples of discrimination techniques include those
88 models based on discriminant analysis (Rogers et al. 1996), general linear model regression
89 (GLM) (Cumming 2000), generalized additive models (GAM) (Leathwick & Whitehead

90 2001), Multivariate Adaptive Regression Spline (MARS) (Muñoz & Felicísimo 2004) and
91 decision-tree based methods (Araújo & Williams 2000). On the other hand, climate envelope
92 techniques (e.g., ANUCLIM, BIOCLIM, DOMAIN, FEM, HABITAT, and Mahalanobis
93 distance) used a classic bioclimatic modelling approach (Hirzel et al. 2002, Anderson et al.
94 2003). Comparisons of modelling techniques are a current issue of research (Jeschke &
95 Strayer 2008).

96 MARS has been applied to Mediterranean species in the past few years for many
97 authors dealing with habitat prediction (Muñoz & Felicísimo 2004; Navarro-Cerrillo et al.
98 2006) reporting accurate results for tree species. Classification algorithms commonly used are
99 CART or TREE NET, often less accurate than MARS (Muñoz & Felicísimo, 2004; Segurado
100 & Araujo 2004). Another group of techniques are generic algorithms based on environmental
101 dissimilarity metrics (BIOCLIM, HABITAT, DOMAIN o ENFA) (Beaumont et al., 2005).
102 More recently, a machine learning approach to predictive modelling has applied new
103 algorithms such as GARP (Genetic Algorithm for Rule-Set Prediction) (Stockwell & Peterson
104 2002, Anderson et al. 2003, Pape & Gaubert 2007) or MAXENT (Maximum Entropy)
105 (Phillips & Dubik 2008), which has proven especially successful in predicting species
106 potential distributions under a wide variety of situations.

107 Modelling prediction methods results in different geographic distributions depending
108 mainly on species ecology specialization and number of samples (Stockwell & Peters 2002;
109 Randin et al. 2006, Pearson 2006), so it is important to compare the performance of the many
110 algorithms and approaches to project distributions within the same regions used to train the
111 models (Elith et al. 2006, Guisan et al. 2007). Most studies have focused on species with a
112 wide geographical distribution and large number of presence data (Beaumont et al. 2005;
113 Phillips & Dubik 2008). On the contrary, spatial distribution modelling at a local scale has

114 been less tested and experience obtained with habitat prediction of spread species cannot be
115 easily generalized.

116 The aim of this contribution is to provide a comparison between modelling algorithms
117 to study the habitat distribution of such a fragmented species as *Buxus balearica* Lam. by
118 evaluating the prediction capability of different spatial distribution models at a local scale.

119 2. Methods

120 2.1. Study area

121 *Buxus balearica* was chosen as the species for analysis based on the criteria of
122 ecological relevance and fairly broad geographic distributions, giving sample sizes sufficient
123 for analysis. The distribution area of *Buxus balearica* studied in this work was the Málaga and
124 Granada populations, which includes almost its total Iberian Peninsula distribution with the
125 exception of the Rágol locality (Almeria province) (36° and 38° north) (Figure 1). This area
126 extends along the southern slope of the Sierras de Almirajara, Cázulas and Los Guájares. Part of
127 these mountains is nowadays protected by the Andalusian Government, the Natural Park of
128 the Sierras de Tejeda, Almirajara and Alhama, with 40,600 ha. The study area was restricted to
129 basic soils, mainly marls and chalky ones, which occupied a surface of 38,180 km².

130 Those mountain ranges form a geographical barrier between the provinces of Málaga
131 and Granada and constitute a continuous alignment of rocky escarpments reaching the most
132 prominent height in the South of Spain and Portuguese coast (Maroma summit, 2,068 m). It
133 acts as a corridor, connecting the Sierras of Málaga with Sierra Nevada, due to its longitudinal
134 distribution. The predominant geology is a consistent limestone, although in some areas there
135 are also Eocene loamy sandstones, limestones, and dolomite marbles. The growing season is
136 about 8-months (March to October), and annual precipitation averages from 450 or 500 mm
137 in coastal areas (Almuñécar, Vélez-Málaga) to more than 1000 mm in some mountainous
138 areas of the northern slope (Alfarnate, Zafarraya), the average values being 600-700 mm in

139 medium height mountainous areas facing the Mediterranean Sea (Cómpeta, Canillas).
140 Dominant climate is a semi-arid Mediterranean (sensu Quézel) below 500 m above sea level
141 and sub-humid Mediterranean (in mountains above 600 and 700 m). Monthly average
142 temperatures oscillate between 11°C and 26°C. In spite of its Mediterranean climate, with a
143 long dry period during the summer, Western warm and humid winds from the Strait of
144 Gibraltar reach these mountains. These sea breezes condense at a certain altitude (generally in
145 the evening) and compensate water deficit.

146 Forest species composition includes many myrtled-shape leaf species well adapted to
147 Mediterranean conditions, such as *Pistacia lentiscus* L., *P. terebinthus* L., *Rhamnus alaternus*
148 L., *Buxus balearica*, *Coriaria myrtifolia* L., *Osyris quadripartite* Salzman ex Decaine,
149 *Cneorum tricoccon* L. and *Quercus coccifera* L. among others. *B. balearica* can only be
150 found in the termo and meso-Mediterranean belts.

151 2.2. Data processing

152 The prediction models were tested based on 17 environmental variables
153 (Environmental Information Network of Andalusia-Consejería de Medio Ambiente) (Table
154 1). All data sets were resampled to 10 m resolution for analysis to reflect the spatial resolution
155 of the occurrence data.

156 The topographical data – aspect (ASP), slope (SLP), Euclidean distance to the closest
157 drainage (EDW), and isolation models (MPI) – were derived from a Land Digital Model (10
158 m resolution) (CMA, 1998). Altitude was not considered because the species appears
159 indistinctly distributed from approx. 0 to 800 m. (Elevation range registers within the study
160 area). Isolation was calculated with SHORTWAVC aml application (Felicísimo et al. 2002).
161 Eight isolation models were calculated, one every 45 days throughout the year, on which total
162 radiation received at the surface of the earth over a period of time was estimated (Table 2).

163 Finally, taking into account the preference of *B. balearica* for areas where ambient humidity
164 may accumulate, Euclidean distance to the closest drainage area was calculated.

165 Meteorological data – winter rainfall (TP₁), spring rainfall (TP₂), summer rainfall
166 (TP₃), autumn rainfall (TP₄), mean temperature of warmest month (ATWM), mean maximum
167 temperature of warmest month (AMTWM), mean temperature of coldest month (ATCM),
168 mean minimum temperature of coldest month (AMTCM), – were obtained from a 28 weather
169 station network (National Institute of Meteorology). The time serial data were 1956–2000.
170 The methodology applied to obtain a 100 m grid of meteorological variables was the
171 interpolation of data from each station. This interpolation was based on multiple regression
172 analysis taking into account the position of each point (coordinates x, y), the altitude
173 (obtained from Land Digital Model), the distance to the sea and orientation weight (maximum
174 value of 4 to SW orientation and minimum value of 0 to NE). The litology (LITO) was
175 obtained from a digital lithological map of the study area (scale: 1:100,000).

176 Current locations of *Buxus balearica* were obtained by field surveys using a GPS
177 (GPSmap® 60CSx), from herbarium specimens (Botanical Gardens-University of Córdoba)
178 and from historical citations. A value of presence (1) of *B. balearica* was assigned to 220
179 points and it was considered when the species presented itself, independently of population
180 density.

181 2.3. Modelling methods

182 Several approaches have been used to approximate the ecological niches of species
183 (Elith et al. 2006). In this study, six methods were compared:

184 a) Multivariate adaptative regression spline, MARS, proposed by Friedman (1991)
185 provides an alternative regression-based method for fitting non-linear responses, using
186 piecewise linear fits rather than smooth functions. MARS uses a stepwise addition/deletion

187 strategy with linear splines and it adopts the generalized cross-validation criterion to add and
188 select basic functions.

189 b) Maximum entropy approach to modelling species' distributions (Maxent) (Phillips
190 et al. 2006, Phillips 2006). Maxent estimates species probability distribution by finding the
191 probability distribution of the maximum entropy, subject to a set of constraints that represent
192 our incomplete information about the target distribution. Mathematical approach of Maxent is
193 based on the estimation of probability function for a species satisfying all the bioclimatic or
194 environmental limits associated while maintaining its distribution and maximum entropy. The
195 Maxent algorithm calculates the distance between the given environmental conditions to each
196 occurrence point and selects the closest distance. In terms of potential distribution, the
197 probability of the presence of a species can be interpreted as the probability of not finding any
198 biophysical limitations to the species' existence; therefore, the greater the entropy of a
199 system, the greater the probability of encountering sites without such limitations (Phillips &
200 Dubik 2008).

201 c) Climate envelope techniques.- The climate envelope modelling approach has its
202 foundations in ecological niche theory. Climate envelopes can be defined as constituting the
203 climatic component of the fundamental ecological niche, or the 'climatic niche'. Some
204 bioclimatic models are based on empirical relationships between observed species
205 distributions and environmental variables (Peterson 2006). Some examples are ANUCLIM,
206 BIOCLIM, DOMAIN, FEM, HABITAT, and Mahalanobis distance. They fit a minimal
207 envelope in a multidimensional climate space and use presence-only instead of
208 presence/absence data. On this study, two models were selected: BIOCLIM and DOMAIN.

209 d) Genetic Algorithm for Rule-set Prediction (GARP) (Stockwell & Noble 1992) is a
210 genetic algorithm that creates ecological niche models for species. The models describe
211 environmental conditions under which the species should be able to maintain populations. For

212 input, GARP uses a set of point localities where the species is known to occur and a set of
213 geographic layers representing the environmental parameters that might limit the species'
214 capability to survive. GARP works in an iterative process of rule selection, testing, and
215 incorporation or rejection: first, a method is chosen from a set of possibilities (logistic
216 regression, bioclimatic rules, range rules, negated range rules), and then is applied to the
217 training data and a rule developed to maximize predictivity (Anderson et al. 2003).

218 e) Support vector machines (SVMs) are a group of supervised learning methods that
219 belong to a family of generalized linear classifiers. They can also be considered a special case
220 of Tikhonov regularization. SVMs can perform binary classification (pattern recognition) and
221 real valued function approximation (regression estimation). A special property of SVMs is
222 that they simultaneously minimize the empirical classification error and maximize the
223 geometric margin; hence they are also known as maximum margin classifiers. Support vector
224 machines map input vectors to a higher dimensional space, where a maximal separating hyper
225 plane is constructed. This hyper plane will attempt to split the input data into two classes. The
226 separating hyper plane is that which maximises the distance between the two classes. The
227 model only depends on a subset of the training data, because the cost function for building the
228 model does not care about training points that lie beyond the margin. Intuitively, this would
229 make the classification correct for testing data that was near, but not identical to the training
230 data.

231 2.4. Model evaluation

232 To assess the agreement between the presence-absence and the predictor's records
233 three statistics were used: the area under curve (AUC), the correlation coefficient (COR) and
234 the maximum Kappa (κ). The AUC ranges between 0 and 1. Cohen's Kappa (κ) is one
235 measurement that can be derived from the confusion matrix. As a validation tool (i.e. when
236 the 'truth' is known), it states the overall accuracy of a prediction once the element of chance

237 has been removed (Liu et al. 2005). Otherwise, Kappa can serve as a tool to assess reliability
238 of prediction in terms of relative agreement. A value near 0 indicates no discrimination
239 (agreement by chance); a value of 1 represents perfect discrimination (agreement); a value of
240 > 0.6 is considered 'good' and >0.8 as 'excellent' (Graham & Hijmans, 2006). Kappa is
241 relatively tolerant to zeros in the confusion matrix and considers both omission and
242 commission errors in one parameter. However, Kappa is unimodally dependent on
243 prevalence, i.e. on the proportion of all presences in the full validation dataset (Allouche et al.
244 2006). This would have been a problem in this project, because the presences of known
245 species were few (training: ~ 150 ; test: ~ 70) compared to the large number of random
246 background points (>1000), which were deemed necessary for a reliable confusion matrix
247 given the large study area. Hence, the TSS (Allouche et al. 2006) was computed instead of
248 Kappa. It is defined as $TSS = \text{sensitivity} + \text{specificity} - 1$. In contrast to Kappa, TSS values
249 can be used to compare prediction performance regardless of both the validation dataset size
250 and the prevalence contained therein, while still featuring the same strengths of Kappa: full
251 consideration of sensitivity, specificity and chance (Allouche et al. 2006). Finally, the
252 concordance among models was evaluated through a correlation analysis.

253 **3. Results**

254 The area under curve (AUC) scores for these models showed a value of 0.73 for
255 Maxent, compared with 0.60 for GARP models, suggesting that GARP models are less
256 predictive than Maxent models (Figure 2). The same results were obtained with the kappa (κ)
257 coefficient and the correlation coefficient (COR) (Figure 2). All the other models tested
258 obtained lower accuracy values. The lowest accuracy was ranked by the Environmental
259 Distance model. The correlation matrix of the different models shows the spatial prediction
260 concordance between models (Table 2). The highest correlations were obtained in two
261 separate groups; first, between the generic algorithms based on environmental dissimilarity

262 metrics. The highest correlations were ranked by the Environmental Distance and the SVM
263 and Bioclimate. The second was between Maxent and MARS, obtaining the highest
264 significance of the correlation

265 Maxent and MARS are different algorithms and follow different computational
266 routines although both produced similar potential distribution models on this study. The
267 relative contributions of the environmental variables were similar as well. Figure 3 shows a
268 heuristic estimate of the relative contributions of the environmental variables to the Maxent
269 and MARS model. To determine the estimate in each iteration of the training algorithm,
270 increase in regularized gain was added to the contribution of the corresponding variable. The
271 highest contributions fell on the climate variables: spring rainfall (TP₂), mean temperature of
272 warmest month (ATWM), and mean temperature of coldest month (ATCM) with proportional
273 weights. The variables with lowest contributions were the terrain isolation models. Figure 4
274 shows the results of the jackknife test of a variable importance. The environmental variable
275 with the highest gain when used in isolation is ATCM, which therefore appears to have the
276 most useful information by itself. The environmental variable that decreases the gain the most
277 when it is omitted is ATCM, which therefore appears to have the largest amount of
278 information that was not present in the other variables.

279 Once Maxent was identified as the most accurate model, a test of sensitivity to the
280 reduction in the number of variables and number of presence input data was performed.
281 Figure 5 shows the curve described by the AUC depending on the number of variables
282 introduced in the analysis. The number of variables introduced in each model is explained in
283 Table 3. Results show a decreasing tendency of the AUC with the reduction of the number of
284 variables. However, the model obtains an accuracy of over 0.90 by applying just 3 variables
285 (TP₂, ATWM, and ATCM). In contrast, the reduction in the number of presence input data
286 shows a greater sensitivity (Figure 6). The maximum number of points taken into the analysis

287 reach high values in the accuracy assessment of both Maxent and MARS. A reduction of 50%
288 in the number of presence data shows a reduction in kappa coefficient around 14.5% for
289 MARS and 6.7% for Maxent. Comparing both models, MARS is more sensitive to the
290 reduction in the number of presence input data.

291 All algorithms produced maps that coincided well with the known distribution of the
292 species (Figure 7), although the GARP prediction tended to be overly extensive, and the
293 Maxent model tended to be somewhat underpredicted. GARP models continued to reconstruct
294 many of the species' known distributions. Maxent models, on the other hand, produced an
295 odd pattern coincident with the input data set at higher probability values reconstructing the
296 on-diagonal quadrants or the off-diagonal quadrants, depending on which were used to train
297 the models (Figure 7).

298 **4. Discussion**

299 *4.1 Model evaluation.*

300 Differences in prediction performance between modelling methods at a local scale is
301 shown on this study. In general terms, results agree with those obtained from the application
302 of the same models to the species distribution at a regional and global scale (Elith et al. 2006,
303 Austin 2007). The generic algorithms based on environmental dissimilarity metrics obtain
304 poor results compared with those obtained by the machine learning approach. The accuracy of
305 Maxent was significantly higher than all the prediction models evaluated. The computational
306 efficiency of each model should be evaluated considering three main aspects: pre- processing
307 required for the input data, processing efficiency of the model itself, post-processing required
308 to acquire statistical and cartographical prediction. Generic algorithm models based as
309 BIOCLIM, DOMAIN, SVC, or GARP, are of a great computational efficiency. Implemented
310 in common free software like the open modeller (<http://openmodeller.sourceforge.net>), and
311 identical input data are used for all of them. The user-friendly interface and the fast

312 computation of the algorithm may prove the usefulness of these models at a regional and
313 global scale. Nevertheless, at a local scale, the results show low accuracy values and poor
314 prediction maps. Maxent and MARS significantly improves all the predictions. Although
315 Maxent share all the computational advantages of generic algorithms, MARS presents some
316 drawbacks. The input data have to be introduced in a specific format (.sav) with the
317 environmental data information extracted from the presence/absence data. This process
318 involves the necessity to process all the information with Arc/info. Once the functions are
319 generated, it is necessary to apply a specific routine programmed in AML to obtain the raster
320 prediction with ArcInfo.

321 In contrast, Maxent not only exhibits a high computational efficiency, it also produces
322 an extensive statistical report of the model. In order to test the sensitivity of the model to
323 different analysis were performed. The first one consisted of the reduction of the number of
324 variables. The high accuracy obtained with just three climate variables shows the strong
325 dependency of the species studied on the climatology conditions. This last result is a very
326 important one from the point of view of the study and assessment of the distribution of *Buxus*
327 *balearica*. The last test performed on the Maxent model was the reduction in the number of
328 presence input data. Compared to the MARS model, the reduction in accuracy was
329 significantly lower with the Maxent model. This last result is also very important insofar as
330 the input data of presence are relatively inaccessible. Apart from this difference in accuracy
331 and computational efficiency, the prediction maps agree between Maxent and MARS, as well
332 as the importance of the variables in the model.

333 4.2. Potential distribution model of *Buxus balearica*.

334 Within the meso-Mediterranean and supra-Mediterranean climate types, *B. balearica*
335 was observed at 220 locations. The digital model developed for predicting the distribution of
336 *B. balearica* is a probabilistic raster map, shown in Figure 7. Kappa index results achieved

337 0.84 accuracy rating in predicting both presence and absence of the species. Response curve
338 diagrams produced along with the model output indicate the effect of individual variables on
339 the Maxent prediction. Environmental variables based on meteorological data made the
340 highest contribution to the potential distribution. The single variable making the greatest
341 contribution to defining the potential distribution of the species was mean temperature of
342 warmest month (ATWM). The species presents an alarming sensitivity to thermic variations.
343 *B. balearica* limits itself to areas with average temperatures ranging from 24.6 to 25.4°C
344 during the warmest month. Results of the mean maximum temperature of warmest month
345 (AMTWM) indicate a species tolerance within a range between 31°C and 32°C. As a
346 consequence, a climate change scenario of solely one degree could greatly diminish the
347 survival possibilities of this species. It is also clear that the low tolerance of *B. balearica* to
348 frost, disappearing as a result of competition with other species in areas where mean
349 temperature of coldest month (ATCM) is below 7°C and mean minimum temperature of
350 coldest month (AMTCM) is below 3°C.

351 Seasonal precipitation also showed itself to have a great impact on the distribution
352 model, measured by total trimester precipitation. Spring rainfall (TP₂) showed yet again the
353 species's sensitivity to climatic and irregularity during the spring-summer season. It could be
354 said that in areas with the presence of *B. balearica* annual precipitation is above 500 mm,
355 with spring precipitations above 110 mm and summer rainfall (TP₃) at least 25 or 30 mm.
356 Consequently, *B. balearica* requires certain soil moisture. However, this interpretation may be
357 incomplete, considering the lack of data concerning crypto-precipitations. This added
358 condensation is continually scarcer due to the climate variations observed in the last decades
359 within the region.

360 The response to orientation models MDI5 (12.5°) and MDI4 (22.5°) with respect to
361 MDI2 (0°) showed a significant contrast. During the summer solstice (sun at its maximum

362 height), the majority of locations with presence of the species receive a great amount of
363 energy while during the winter solstice (sun at its lowest position) solar radiation is at its
364 lowest. According to model results, *B. balearica* distribution is more affected by solar
365 isolation during the period between the summer solstice and fall equinox (MDI5-12.5°). The
366 distribution of the species, on the other hand, was indifferent to aspect, even though a slight
367 preference was shown for a southerly orientation. These results agree with field observations,
368 in which *B. balearica* was seen to be distributed with various orientations. Slope was also no
369 significant in determining the presence or absence of the species, as it appears in slopes of a
370 varying steepness.

371 Euclidean distance to nearest drainage, based on a watershed model, was an influential
372 factor in determining the presence of the species. *B. balearica* specimens appear to take
373 refuge in precipices, streams and rivers, and certainly not so much because of subsoil
374 moisture, but more likely because it is in such areas that the species is subject to less incident
375 radiation. Even more so because in these areas it is able to avoid forest fires, which here do
376 not reach great intensity, and conditions allow the plant resprouting. Moreover, seedling
377 mortality due to summer drought can be overcome in the riparian zone (brooks and streams of
378 seasonal flows).

379 **5. Conclusion**

380 An evaluation of model performance to predict the distribution of *Buxus balearica* is
381 described in this paper. It is important to note that many applications of species distribution
382 models depend on predicting potential distributions, rather than realized distributions. Facing
383 increasing climate change and human pressure, *B. balearica* populations will tend to shelter in
384 areas with microclimate conditions that are less unfavourable and which limit
385 evapotranspiration, regardless of altitude or topography. Average performance of different
386 habitat distribution models for current versions showed that Maxent, evaluated on

387 independent presence/absence test, may be applied to predict potential distributions of *B.*
388 *balearica*.

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396

For Review Only

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497 **Table 1.** Environmental variables resampled to 10 m resolution used on the prediction
 498 models.

Variable	COD	Source
Aspect	ASP	Generated from DEM
Slope	SLO	Generated from DEM
Euclidean distance to watershed	EDW	Generated from DEM
Model of Potential Insolation , Solar Declination 22.5°	MPI+22.5	Generated from DEM
Model of Potential Insolation, Solar Declination 12.5°	MPI+12.5	Generated from DEM
Model of Potential Insolation, Solar Declination 0°	MPI 0	Generated from DEM
Model of Potential Insolation, Solar Declination -12.5°	MPI-12.5	Generated from DEM
Model of Potential Insolation, Solar Declination -22.5°	MPI-22.5	Generated from DEM
Total Precipitation (January- March)	TP1	National Institute of metheroology (INM) (www.mapya.es)
Total Precipitation (April-June)	TP2	National Institute of metheroology (INM) (www.mapya.es)
Total Precipitation (July- September)	TP3	National Institute of metheroology (INM) (www.mapya.es)
Total Precipitation (October- December)	TP4	National Institute of metheroology (INM) (www.mapya.es)
Average temperature of the warmest month	ATWM	National Institute of metheroology (INM) (www.mapya.es)
Average maximum temperature of the warmest month	AMTWM	National Institute of metheroology (INM) (www.mapya.es)
Average temperature of the coldest month	ATCM	National Institute of metheroology (INM) (www.mapya.es)
Average minimum temperature of the coldest month	AMTCM	National Institute of metheroology (INM) (www.mapya.es)
Lithology	LITO	Department of Environment of the Government of Andalusia

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500 **Table 2.** Bivariate Pearson Correlation matrix of the concordance of the different spatial
 501 distribution models analysed.

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	Bioclim	Maxent	MARS	GARP	SVM	ENV_DIST
Bioclim	1.000					
Maxent	-0.040	1.000				
MARS	-0.180	0.807***	1.000			
GARP	0.769**	0.303*	0.240	1.000		
SVM	0.856**	0.290	-0.110	0.798**	1.000	
ENV_DIST	0.856**	0.290	-0.110	0.798**	1.000**	1.000

503 **. Correlation is significant at the 0.01 level (2-tailed)

504 ***. Correlation is significant at the 0.001 level (2-tailed)

505 **Table 3.** Sensitivity analysis of Maxent model. The table shows from right to left, the
 506 variables introduced, the code assigned to the model and the relative importance of the last
 507 variable introduced. Variables on table 1.

508

Model	Cod	Imp.
ATWM	Mod_01	24
ATWM, ATCM	Mod_02	12.9
ATWM, ATCM, TP2	Mod_03	12.2
ATWM, ATCM, TP2, AMTCM	Mod_04	7.8
ATWM, ATCM, TP2, AMTCM, MPI-22.5	Mod_05	5.8
ATWM, ATCM, TP2, AMTCM, MPI-22.5	Mod_06	5.6
ATWM, ATCM, TP2, AMTCM, MPI-22.5, TP3	Mod_07	5.3
ATWM, ATCM, TP2, AMTCM, MPI-22.5, TP3,	Mod_08	5.2
ATWM, ATCM, TP2, AMTCM, MPI-22.5, TP3, TP4	Mod_09	5.1
ATWM, ATCM, TP2, AMTCM, MPI-22.5, TP3, TP4, TP	Mod_10	4.7
ATWM, ATCM, TP2, AMTCM, MPI-22.5, TP3, TP4, TP, AMTWM	Mod_11	3.8
ATWM, ATCM, TP2, AMTCM, MPI-22.5, TP3, TP4, TP, AMTWM, EDW	Mod_12	3.5
ATWM, ATCM, TP2, AMTCM, MPI-22.5, TP3, TP4, TP, AMTWM, EDW, TP1	Mod_13	2.1
ATWM, ATCM, TP2, AMTCM, MPI-22.5, TP3, TP4, TP, AMTWM, EDW, TP1, MPI0	Mod_14	0.7
ATWM, ATCM, TP2, AMTCM, MPI-22.5, TP3, TP4, TP, AMTWM, EDW, TP1, MPI0, MPI-12.5	Mod_15	0.7
ATWM, ATCM, TP2, AMTCM, MPI-22.5, TP3, TP4, TP, AMTWM, EDW, TP1, MPI0, MPI-12.5, ASP	Mod_16	0.4
ATWM, ATCM, TP2, AMTCM, MPI-22.5, TP3, TP4, TP, AMTWM, EDW, TP1, MPI0, MPI-12.5, ASP, MPI+12.5	Mod_17	0.1
ATWM, ATCM, TP2, AMTCM, MPI-22.5, TP3, TP4, TP, AMTWM, EDW, TP1, MPI0, MPI-12.5, ASP, MPI+12.5, MPI+22.5	Mod_18	0.1

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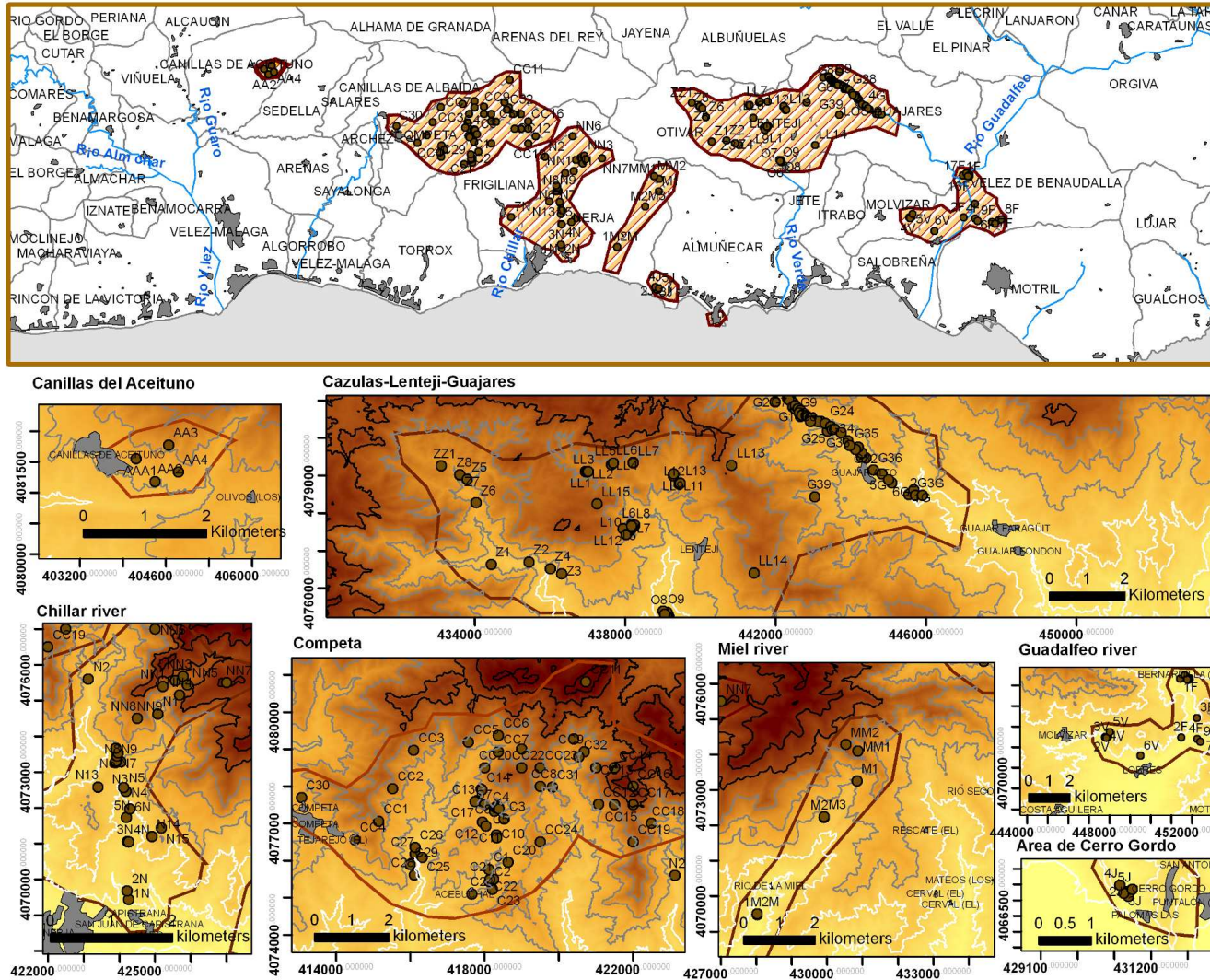
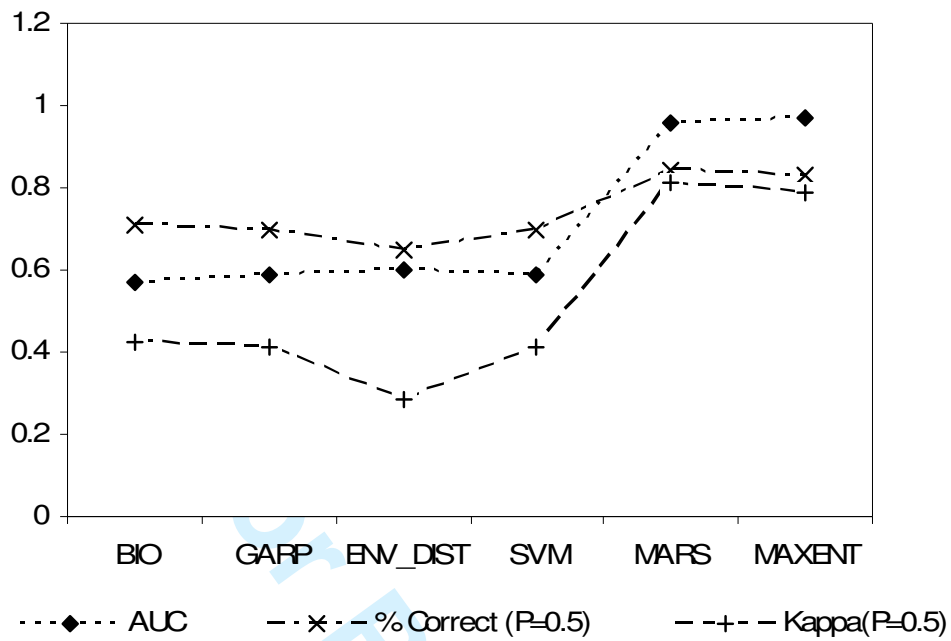


Figure 1. Distribution of *Buxus balearica* on the study locations area (Southern Spain).

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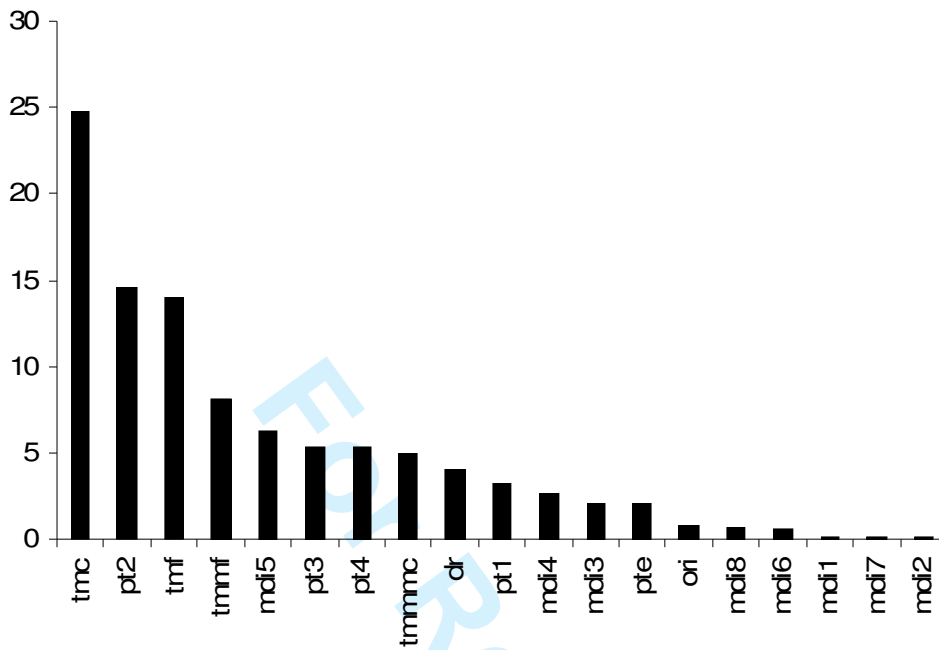
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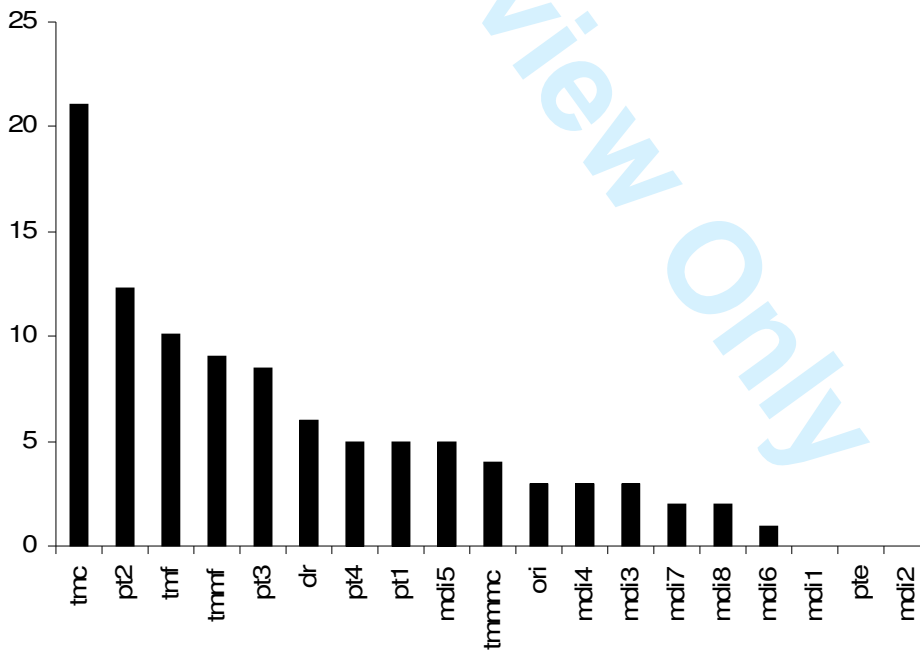
534 **Figure 2.** Comparison of the validation coefficients obtained by the application of six
 535 different spatial distribution models: BIOCLIM (BIO), GARP, DOMAIN (ENV_DIST),
 536 SVM, MARS and MAXENT.

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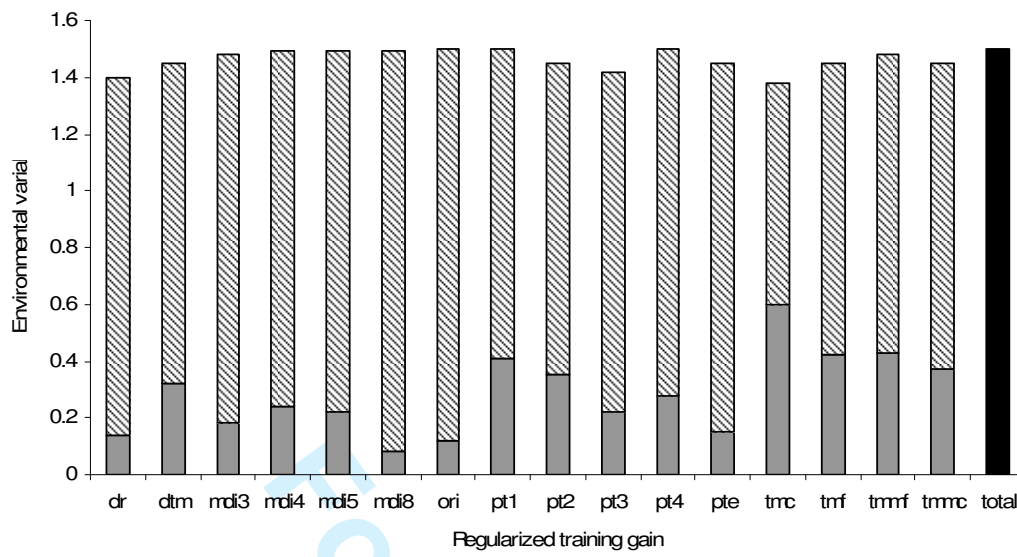
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Figure 3.- Relative contributions of the environmental variables to the Maxent model (a) and to the MARS model (b).

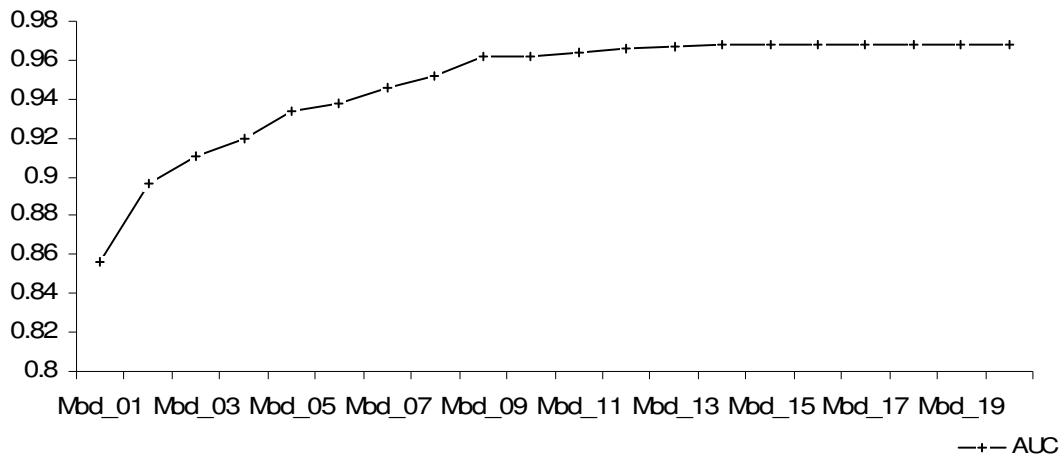
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Figure 4.- Jack-knife test of variable importance to the Maxent model.

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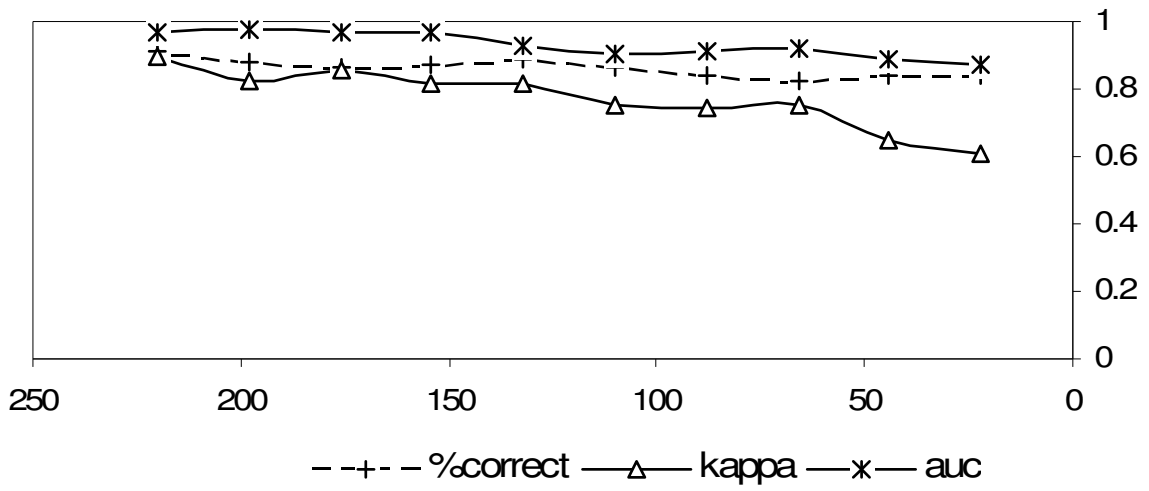


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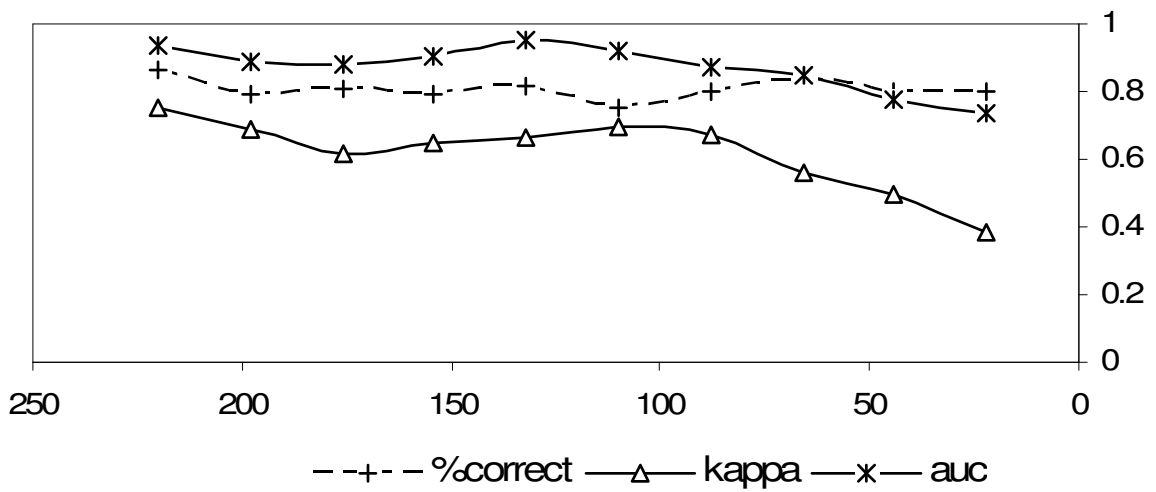
Figure 5.- Sensitivity analysis of Maxent model based on the reduction in the number of variables.

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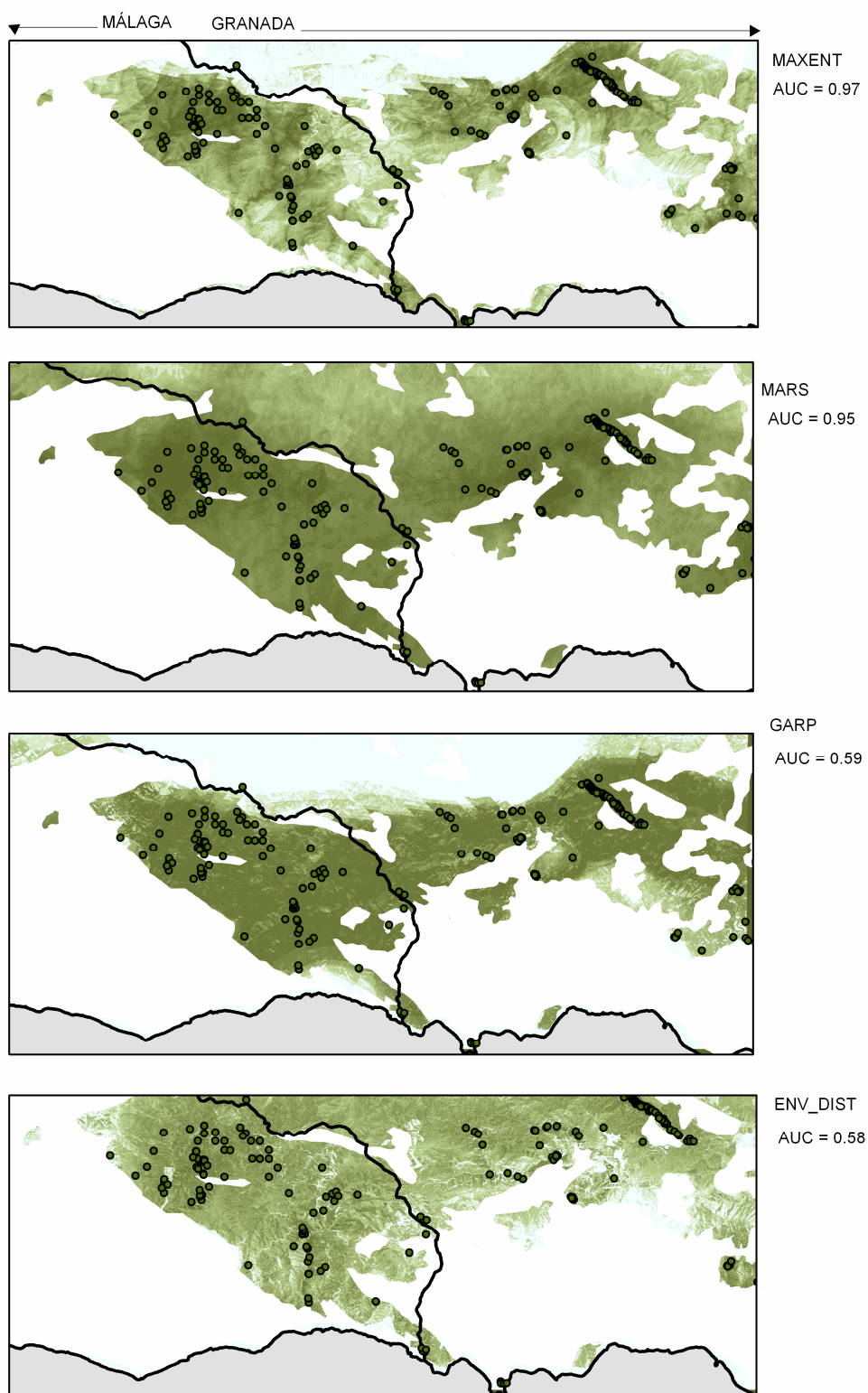


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571 **Figure 6.-** Accuracy functions of Maxent model and MARS depending on the number of
572 presence input data.
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578 **Figure 7.-** Potential distributions of *Buxus balearica* obtained by the application of MAXENT,
579 MARS, GARP and ENVIRONMENTAL DISTANCE (ENV_DIST).

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